



2009-01-01

Land use Planning in Ireland: a Life Cycle Energy Analysis of Recent Residential Development in the Greater Dublin Area

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Duffy, A.: Land use planning in Ireland: a life cycle energy analysis of recent residential development in the Greater Dublin area. *International Journal of Life Cycle Assessment*, Vol.14, 3, 2009, pp.257-267. doi:10.1007/s11367-009-0059-7

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Land use planning in Ireland—a life cycle energy analysis of recent residential development in the Greater Dublin Area

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Received: 26 August 2008 / Accepted: 6 February 2009
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Abstract

Background, aim, and scope One third of the total housing stock in the Republic of Ireland has been built in 10 years up to and including 2006 and of this approximately 34% was built in the Greater Dublin Area (GDA). Much of the housing was low-density with poor public transport links leading to doubts over its sustainability—particularly in terms of energy use. Although the country is committed to reducing greenhouse gases to 13% above 1990 levels by the period 2008–2012, by 2005, emissions were already 25.4% higher than the baseline and current projections are that this figure will rise to 37% over the period. The residential sector is estimated to contribute to approximately 24.5% of energy-related CO₂ emissions. This paper estimates total emissions from residential developments in the GDA constructed between 1997 and 2006.

Materials and methods Carbon dioxide equivalent (CO₂) emissions are estimated using a life cycle assessment approach over a 100-year building lifespan and employing process, input–output and hybrid energy techniques. Life cycle stages include: construction, operation, transport, maintenance and demolition. The main data sources include: national population and industry census data, household travel survey data, residential energy performance surveys and national accounts. The GDA was split into four zones each encompassing development at increas-

ing radii from Dublin's city centre, namely: city centre, suburbs, exurbs and commuter towns.

Results Per capita CO₂ life cycle emissions in the GDA were found to be approximately 50–55% greater in the exurbs and commuter towns than in the city centre. Of the five life cycle stages studied, operational energy requirements (predominantly space heating and hot water, but including power) contributed most significantly to emissions (68%), followed by transport (17%), construction (9%) and maintenance/renovation (6%).

Discussion Operating emissions from dwellings in the commuter town and extra-urban zones were almost twice those in the city centre both due to larger dwelling sizes and the predominance of detached and semi-detached dwellings (with large amounts of exposed walls) in the former and the prevalence of smaller apartments in the latter. Car use was most pronounced in the zones furthest from the city centre where per capita emissions were almost twice those of residents in the city centre. Despite their smaller size, the per capita construction CO₂ emissions of apartments were approximately one third greater than for low-rise dwellings due to the greater energy intensity of the structure. However, this difference was more than compensated for by the significantly lower operational emissions referred to above.

Conclusions In 2006, recurrent CO₂ emissions (operational, transport and maintenance) from dwellings built in the GDA over the ten preceding years were 2,108 kt while construction-related emissions in that year were 1,325 kt giving a total contribution from the residential sector of 3,434 kt CO₂/annum—representing 4.9% of national emissions for that year. Had the development policy prescribed 'city centre'-type development and transport modes, then emissions for the year 2006 would have been 2,892 kt CO₂—a reduction of almost 16% over the actual figure. However, in this scenario recurrent emissions would have

Responsible editor: Llorenç Milà i Canals

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71 been reduced to 1,417 kt CO₂—a reduction of 33% over
 72 actual levels.
 73 *Recommendations and perspectives* This study supports
 74 Irish and international governments' policies aimed at
 75 curbing CO₂ emissions from the domestic sector which
 76 focus primarily on reducing operational emissions from
 77 new and existing housing through design and construction
 78 improvements. However, it demonstrates that significant
 79 reductions in operational emissions are associated with
 80 high-density residential development with modest floor
 81 areas. Furthermore, it highlights the scope for transport
 82 emissions' reductions through better spatial planning
 83 leading to reduced car travel.

84 **Keywords** Carbon dioxide equivalent · CO₂ · Domestic
 85 dwellings · Embodied energy · Energy · Greenhouse gas
 86 emissions · Life cycle assessment · Spatial planning

87 **1 Background, aim, and scope**

88 At the end of 2006 there were 1,835,515 domestic dwell-
 89 ings in the Republic of Ireland (CSO 2007a). Of these,
 90 607,961—representing one third of the total housing
 91 stock—were built in the 10 years up to and including that
 92 year (Fig. 1). Approximately 34% of this new housing was
 93 built in Dublin and the greater Dublin region, much of
 94 which was low-density with poor public transport links.
 95 Indeed, much of the housing development in Ireland can be
 96 characterised as 'once-off' or low-density located outside
 97 urban centres: over one third of the development comprised
 98 detached houses, 44% were semi-detached or terraced
 99 housing units and only one fifth were apartments. In
 100 contrast, multi-family dwellings account for almost half of
 101 Europe's housing stock (Netherlands Ministry of Housing,
 102 Spatial Planning and the Environment 2004). The adoption

in Ireland of such low density suburban and extra-urban
 development policies may have resulted in increased
 greenhouse gas emissions when compared with higher
 density urban developments.

Ireland's energy emissions' performance is poor by
 international standards: in 2004, energy use per capita was
 3,870 kg of oil equivalent (kgoe) and associated greenhouse
 gas emissions were 10,589 kg. This compares unfavourably
 with EU27 figures of 3,689 kgoe and 8,180 kg, respectively
 (Eurostat 2007). Although the country has committed to
 reducing greenhouse gases to 13% above 1990 levels by
 the period 2008–2012, by 2005, emissions were already
 25.4% higher than the baseline (EPA 2007) and current
 projections are that this figure will rise to 37% over the
 period. Residential energy use accounted for almost one
 quarter of total national fuel consumption in 2005 of which
 almost 75% is used for space heating and hot water
 (Howley et al. 2006).

The impact of a domestic development on greenhouse
 gas emissions is related to the quantity and carbon content
 of the energy required over its lifetime. For example, two
 similar buildings with different passive thermal perform-
 ances will consume different amounts of energy: one will
 emit more greenhouse gases than the other. Identical
 buildings powered by fuels with different carbon contents
 (for example, biomass- versus oil-fuelled boilers) will also
 have differing impacts. The energy impact of a domestic
 development over its entire life cycle can be viewed as the
 sum of this operational energy use together with the
 energy needed to produce and maintain it, demolish it as
 well as the energy required to travel to and from it.
 Similarly, the CO₂ impact is the global warming potential-
 weighted sum of greenhouses gases (predominantly
 carbon dioxide, nitrous oxide and methane) emitted by
 this energy use.

Australian literature suggests that the average energy
 required to produce a house in that country is approximately
 5 GJ/m² producing some 0.49 tonnes/m² floor area of carbon
 dioxide (CSIRO 2007) based on Australian emissions'
 intensities. This is referred to as 'embodied energy' and
 theoretically includes all direct and indirect production
 energy inputs such as construction, material production,
 raw material extraction and associated services. The
 average floor area of a domestic dwelling in Ireland in
 2004 was 112 m² (SEI 2005) and applying the Australian
 data, this gives a per-dwelling embodied energy of
 560 GJ and emissions of 55 tonnes of carbon dioxide
 (CO₂). Assuming a 100-year building lifespan, this is
 equal to annualised emissions of 0.55 tonnes of CO₂ and
 5.6 GJ of energy use. Although these figures are rough
 estimates—since the Australian embodied energies are
 likely to differ from those in Ireland (no Irish data are
 available at the time of writing)—when used with the

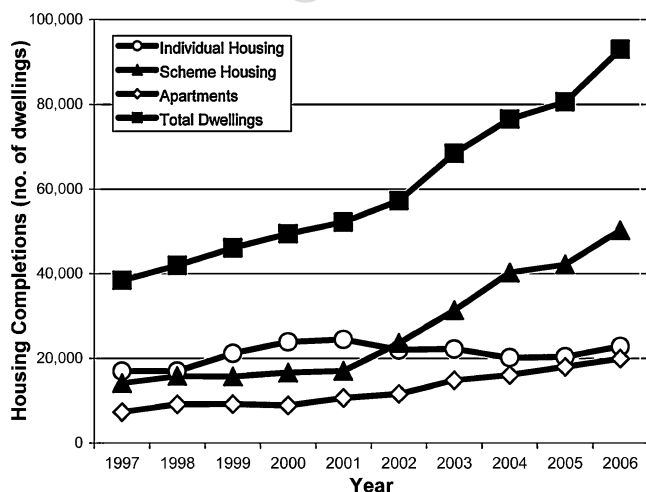


Fig. 1 Annual housing completions in Ireland 1997–2006

156 operational and travel data below, they indicate the
 157 relative importance of embodied energy in the life cycle
 158 of a domestic building and will assist in the development
 159 of the life cycle assessment (LCA) methodology described
 160 in this paper.

161 In 2004, the mean energy use per Irish dwelling was
 162 88 GJ resulting in the emission of 8.2 tonnes of CO₂ per
 163 annum. Some 79% of energy use was in the form of direct
 164 fossil fuel consumption, the remainder was electricity (SEI
 165 2005).

166 In 2006, the Irish transport sector consumed 213,000 TJ
 167 of energy and emitted 15,273 kt of CO₂ (SEI 2006) of
 168 which 40% was attributable to private cars (SEI 2004). In
 169 the same year, there were 1,469,521 households in Ireland
 170 (CSO 2007b) giving average annual transport-related CO₂
 171 and energy use figures of 4.2 tonnes and 58 GJ per
 172 dwelling, respectively. This approximation overestimates
 173 the direct energy inputs associated with moving to and
 174 from a domestic dwelling since many private car trips
 175 would be used for other purposes, although it ignores
 176 indirect energy inputs such as product development and
 177 manufacturing.

178 Little data are available on the maintenance-related
 179 energy requirements of domestic dwellings and associated
 180 emissions. Figures for the refurbishment and repair of office
 181 buildings are reported to lie in the range of 0.17 to 0.34 GJ/
 182 m² (Yohanis and Norton 2002). Taking the lower figure—
 183 since office refurbishment is likely to be more energy-
 184 intensive than domestic refurbishment—and assuming an
 185 interval between refurbishments of 7 years, the annual
 186 maintenance-related energy intensity for an average Irish
 187 house is 2.72 GJ/m² with associated emissions of 0.23 t
 188 CO₂. However, even the use of the lower office refurbish-
 189 ment energy intensity figure may overestimate domestic
 190 emissions.

191 Based on these data, the total approximate annualised
 192 life cycle energy requirement for an average domestic Irish
 193 dwelling is 154.3 GJ emitting 13.2 tonnes of CO₂.
 194 Transport and operation dominate life cycle energy require-
 195 ments (38% and 56% respectively) and CO₂ emissions
 196 (32% and 62% respectively). Based on these preliminary
 197 results, the methodology outlined in this paper focuses
 198 therefore to a greater extent on operational and transport
 199 energy use and emissions than on construction, mainte-
 200 nance and demolition.

201 The objectives of this study are to: determine the whole-
 202 life energy and CO₂ intensities of dwellings built in the
 203 Greater Dublin Area (GDA) between 1997 and 2006; assess
 204 the relative impact of different life cycle stages on energy
 205 use and CO₂ emissions; and quantify and compare the
 206 impact of urban, suburban and extra-urban residential
 207 developments in and around Dublin on national energy-
 208 related greenhouse gas emissions.

2 Materials and methods	209
Life cycle assessment (LCA) was used to determine energy use and carbon dioxide emissions for all stages of a residential development including:	210
	211
	212
• construction;	213
• operation including heating, hot water, lighting and small power loads;	214
	215
• travel to and from the development;	216
• maintenance; and	217
• demolition	218
The study only considered the greenhouse gas impacts associated with energy use and ignored wider environmental impacts such as resource depletion, groundwater pollution and habitat loss. CO ₂ emissions only were calculated since these data are easily available. However, since CO ₂ accounts for only 93% of non-agricultural greenhouse gas (GHG) emissions in Ireland by global warming potential (CSO, 2007c), resulting figures were divided by 0.93 to correct for unaccounted energy-related GHG emissions (such as N ₂ O and CH ₄) to estimate total emissions in CO ₂ equivalent (referred to hereafter simply as CO ₂).	219
	220
	221
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	230
Energy requirements and emissions for each of the life stages of the residential development were determined using process, input–output or hybrid process analysis depending on data availability. These methodologies are described by inter alia Crawford (2007) and Bullard et al. (1978).	231
	232
	233
	234
	235
	236
2.1 Developments	237
Two developments were selected to represent typical recent residential developments in the Greater Dublin Area (a term which is used to describe the city and county of Dublin as well as the adjacent counties of Kildare, Meath and Wicklow) which are described below.	238
	239
	240
	241
	242
A high density mixed-use urban development in the city of Dublin comprising 300 apartments of one, two and three bedrooms with a total gross floor area of 22,500 m ² . The scheme also contains retail and office units.	243
	244
	245
	246
A development of 118 two-storey detached, semi-detached and terraced houses on a 5.7 ha site outside the city comprising detached, semi-detached and terraced houses ranging in size from three to five bedrooms.	247
	248
	249
	250
2.2 Construction	251
The energy embodied in the above developments was determined using hybrid process analysis up to the point of completion of the construction process by the contractor and hand over to the new resident. This required material-	252
	253
	254
	255

256 specific energy intensities (Hammond and Jones 2006)
 257 which were used in conjunction with national input–output
 258 tables (CSO 2006a) to quantify the total upstream and
 259 direct energy requirements for the building. Upstream
 260 energy requirements might include raw materials' extrac-
 261 tion and building materials' manufacturing; direct energy
 262 relates to energy expended during the construction process
 263 itself.

264 Bills of quantities detailing material quantities and prices
 265 were obtained from the contractor who built the schemes.
 266 Where possible, materials were identified, characterised and
 267 process energy intensities applied. Energy intensities of
 268 unidentified expenditure were estimated using Irish con-
 269 struction sector energy intensities derived from national
 270 input–output tables, mean national energy tariffs, primary
 271 energy factors and disaggregation coefficients (Acquaye
 272 et al. 2008). An integral part of this methodology involved
 273 calculating a carbon dioxide equivalent intensity based on
 274 the mix of fuels used.

275 **2.3 Transport**

276 Transport emissions for all trips to and from dwellings were
 277 calculated based on data from the Dublin Transportation
 278 Office's 2006 Household Survey (DTO 2006) of approxi-
 279 mately 2,500 households in the Greater Dublin Area. The
 280 survey was carried out over a 1-week period where
 281 participants kept a detailed log of all movements to and
 282 from their dwellings. Relevant variables were extracted
 283 from the database which included:

- 284 • mode of transport (for example train, rapid rail, tram,
 285 bus, truck, van, car, taxi, motorbike, running, walking);
- 286 • number of trips made;
- 287 • trip distances; and
- 288 • trip duration

289 It is accepted that development density and land-use mix
 290 are inversely related to household-related travel distances
 291 and vehicle emissions (Frank et al. 2000; Næss 2006;
 292 Newman and Kenworthy 1999). Moreover, the distance of
 293 a residence from concentrated employment and recreational
 294 areas (typically a city centre) affects both transport mode
 295 (DTO 2006) and travel distance and, therefore, fuel
 296 consumption and CO₂ emissions. In order to capture the
 297 effects of density and distance to urban centres, the Greater
 298 Dublin Area Household Survey data were therefore divided
 299 into four zones of increasing distance from the city centre
 300 (Fig. 2). These are:

302 **Zone 1** City Centre which is roughly bounded by two
 303 canals and the inner ring roads to the north and
 304 south giving a radius of up to 3.0 km around the
 305 city;

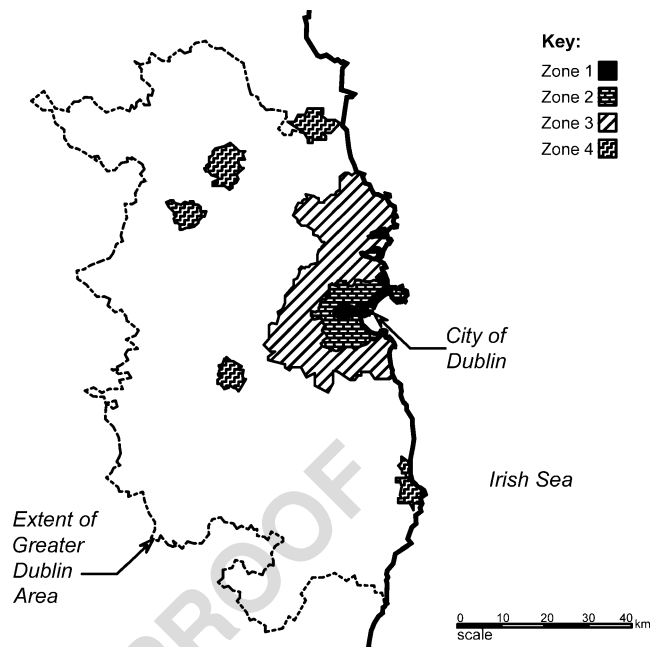


Fig. 2 Map of the Greater Dublin Area (GDA) showing the zones analysed

- 306 **Zone 2** Suburbs which are located between the City
 307 Centre and the M50 motorway ring road at a
 308 radius of approximately 9.0 km around the city;
- 309 **Zone 3** Exurbs covering the low density urban areas in the
 310 remainder of county Dublin located outside the
 311 M50 incorporating many of the closer commuter
 312 towns at a radius of 15 to 30 km from the centre;
 313 and
- 314 **Zone 4** Commuter towns of Drogheda (42 km from the
 315 centre), Navan (45 km), Enfield (40 km), Naas
 316 (31 km) and Wicklow (43 km).

317
 318 Data were then analysed by zone to determine the total
 319 number of trips by mode of transport and total annualised
 320 modal distances travelled. Mean data were then calculated
 321 on a per household and per capita basis.

322 Fuel consumption and CO₂ emissions data were calcu-
 323 lated in the following ways:

324 no carbon dioxide emissions or fuel consumption was
 325 attributed to walking, running or cycling. This assumes
 326 that the energy and CO₂ embodied in equipment such
 327 as footwear, bicycles, special clothing and helmets is
 328 negligible compared to that embodied in other forms of
 329 transport. The food requirements of individuals are
 330 outside the LCA boundary in all cases.
 331 aeroplane, train, rapid rail transit, tram, bus, mini-bus,
 332 truck and motorbike energy uses and emissions were
 333 calculated using published average emissions per
 334 person kilometre (Cox and Hickman 1998; Davies

335 and Diegel 2007). Trains, buses, mini-buses and trucks
 336 were assumed to be diesel-fuelled while DART (a rapid
 337 rail system) and LUAS (a tram system) use electricity
 338 only. The average fuel consumption and carbon
 339 dioxide emissions data used in the analysis is presented
 340 in Table 1.

341 Since emissions from cars were expected to predomi-
 342 nate, these were calculated in the more detailed manner
 343 described below.

344 Engine size and fuel type (diesel/petrol) distributions
 345 were determined using 2005 local government car registra-
 346 tion data and a weighted mean engine size was calculated
 347 for both fuel types.

348 Mean fuel consumption was calculated for this engine
 349 size based on published data and speed-related efficiencies
 350 (Davies and Diegel 2007) were combined to give an
 351 equation for fuel consumptions for the weighted mean
 352 engine size (calculated above) at different speeds.

353 Average speeds for all car trips were calculated and fuel
 354 consumption determined using the fuel consumption to
 355 speed relationship.

356 This represents the fuel consumed only and does not
 357 consider the full life-cycle energy use of the car. Research
 358 by Castro et al. (2003) suggests that direct fuel consump-
 359 tion represents approximately 95% of the total life cycle
 360 energy requirements of a car and the results were adjusted
 361 accordingly. In the absence of equivalent life cycle
 362 assessment literature for other modes of transport, this
 363 figure was also applied to buses, trains, vans, trucks and
 364 motorbikes.

365 **2.4 Operation**

366 Operational energy requirements were quantified using the
 367 Energy Performance Survey of Irish Housing undertaken
 368 jointly by Sustainable Energy Ireland and CODEMA
 369 between 2004 and 2005. This study involved a detailed
 370 survey of the energy use characteristics of dwellings in
 371 Ireland and covered a range of construction types, ages and
 372 occupancy patterns. Relevant survey data were selected to
 373 be representative of the development types chosen here, in
 374 particular:

375 Data in the fields ‘Detached’, ‘Semi-Detached’,
 376 ‘Terraced’ and ‘Purpose-Built Apartment’ were incor-

porated whereas, ‘Converted Apartment’ and ‘Other’ 377
 were ignored and 378
 Developments built since 1997 were included since 379
 different Building Regulations and energy performance 380
 standards pertained prior to this date and would not 381
 have been representative of the 1997–2006 housing 382
 stock. However, a lack of data for apartments neces- 383
 sitated the use of pre-1997 data which would have 384
 overestimated energy use and emissions to some 385
 extent. 386

Mean annualised energy use and carbon dioxide emis- 387
 sions data were calculated for the reduced sample of 59 388
 dwellings based on recorded fuel mixes and quantities and 389
 these were validated against Scottish data. Data were 390
 converted from per-dwelling to per capita by dividing by 391
 the average number of persons per private household for 392
 the county or city as listed in the 2006 national census 393
 (CSO 2007a). 394
 395

2.5 Maintenance and demolition 396

Maintenance requirements were determined using input– 397
 output analysis. Total expenditure for Irish housing repair, 398
 maintenance and improvements for the year 2004 were 399
 used (CSO 2006b) together with the energy intensity for the 400
 construction sector and total existing housing units for the 401
 period to determine the average annual domestic dwelling 402
 maintenance energy use. 403

Demolition costs were calculated using input–output 404
 analysis employing data from the 2004 Census of Building 405
 Construction (CSO 2006c). Turnover and input figures for 406
 Nace 45.1 ‘Site preparation, demolition and wrecking of 407
 buildings, earth moving, test drilling and boring’ were used 408
 to determine energy intensities and these were used with 409
 2004 published demolition costs (Davis Langdon and 410
 Everest 2004) to determine demolition energy requirements. 411

3 Results and discussion 412

3.1 Construction 413

Figure 3 shows the embodied CO₂ emissions per dwelling. 414
 It can be seen that detached dwellings were found to have 415

t1.1 **Table 1** Average fuel consump-
 tions and CO₂ emissions for
 various modes of transport ex-
 cluding cars (source: EEA 2003)

Mode	Fuel consumption (grams oil equivalent/passenger km)	CO ₂ emissions (g/passenger km)	t1.2
Bus	23	73	t1.3
Truck	97	216	t1.4
Motorbike	36	104	t1.5
Rail	20	44	t1.6

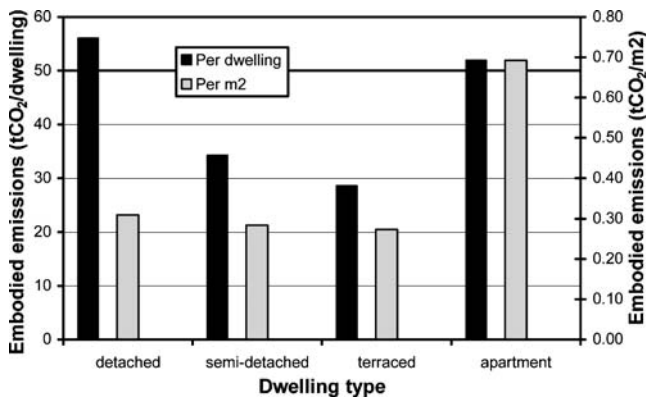


Fig. 3 Embodied CO₂ for each dwelling type studied expressed both on a per dwelling and a per m² of gross floor area basis

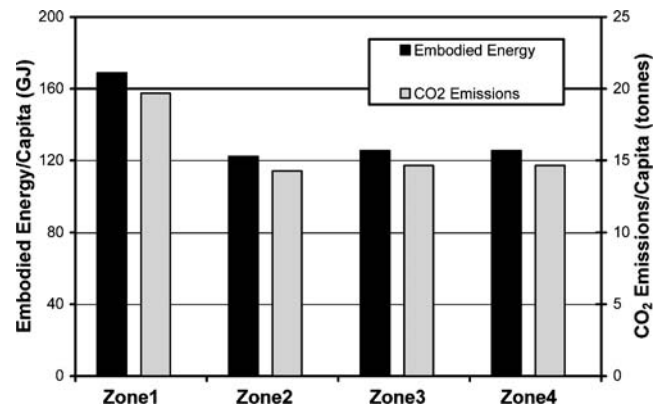


Fig. 5 Per capita embodied energy and construction-related CO₂ emissions for each zone

416 the highest embodied emissions due to their large size and
 417 construction characteristics. For example, the mean size of
 418 a detached house in the sample was 181.2 m² whereas
 419 semi-detached houses, terraced houses and apartments were
 420 120.6, 104.6 and 75.0 m², respectively. Detached houses
 421 have greater construction materials' requirements since they
 422 have no shared walls, unlike other dwelling types. Apart-
 423 ments have the second highest embodied energy due to the
 424 need for more structural elements and, consequently,
 425 greater quantities of steel and concrete—both materials
 426 which have relatively high energy intensities. The use of
 427 these materials results in apartments producing the most
 428 CO₂ emissions per constructed dwelling. Terraced houses
 429 were found to have the lowest embodied energy and CO₂
 430 emissions due to their relatively small size and use of
 431 shared structure (such as party walls). When compared on a
 432 per square metre of constructed gross floor area, it can be
 433 seen that due to their relatively small size and large
 434 structural requirements, apartments have almost two and a
 435 half times more embodied energy and CO₂ emissions than
 436 other forms of residential construction.

437 Figure 4 shows the percentage of house type completed
 438 in 2006 by zone where it can be seen that areas closest to

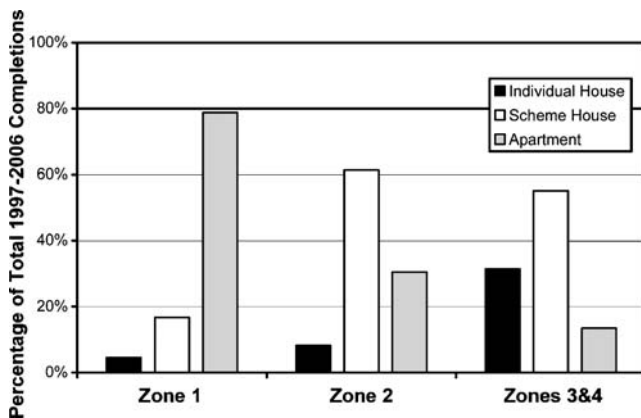


Fig. 4 Percentage of housing completions by type for each zone

439 urban centres had the highest concentration of apartment
 440 completions: almost four fifths of dwellings completed in
 441 the city centre were apartments while this figure fell to 16%
 442 in Zones 3 and 4 where the development of 'scheme
 443 houses' comprising mixed housing units dominated.

444 Based on the mix of dwellings completed in the 10-year
 445 period between 1997 and 2006 and on the average number
 446 of persons per household (CSO 2007a), average embodied
 447 energies and CO₂ emissions were calculated for each zone
 448 and are shown in Fig. 5. It can be seen that both embodied
 449 energy requirements and CO₂ emissions are highest in Zone
 450 1 due to its high proportion of energy-intensive apartments
 451 and relatively low occupancy figures (2.5 persons per
 452 household compared to 3.0 for Zone 4).

3.2 Transport

454 It was found that the further a development is from the city
 455 centre, the greater the reliance on the private car while less
 456 trips are made using public transport. In Zone 1 some 31%
 457 of trips are made by car compared to 49%, 58% and 63%
 458 for Zones 2, 3 and 4, respectively. The percentage of trips
 459 made by walking or cycling decreases with increasing
 460 distance from the city centre: for Zones 1, 2 and 3 the

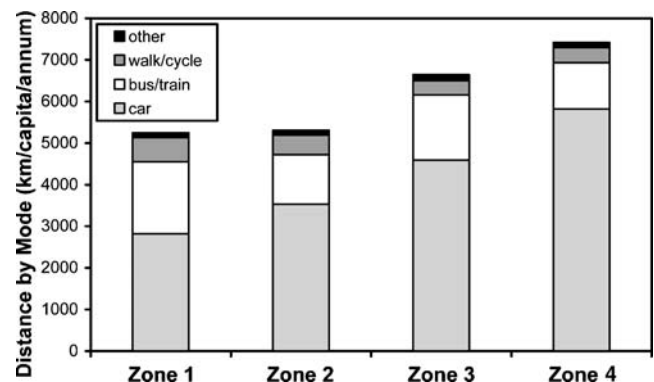


Fig. 6 Annual per capita distance travelled by mode for each zone

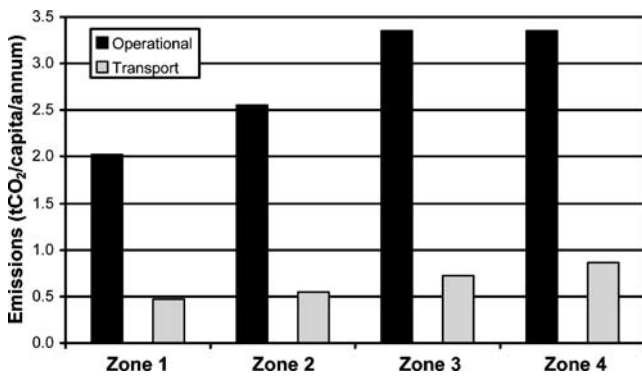


Fig. 7 Annual per-capita operational and transport-related CO₂ emissions for each zone

percentages of trips made using these modes are 40%, 29% and 22%, respectively. However, for Zone 4—commuter towns—the figure increases to 28% suggesting that local employment centres and amenities situated close to residential areas create greater opportunities for walking or cycling than in the Exurbs (Zone 3).

Figure 6 shows the annual modal distances travelled per capita for each zone. There is a clear positive relationship between distance from the centre of Dublin and distance travelled by car. The per capita distances travelled by public transport (bus and train) were highest for Zones 1 and 3 at 1,735 km and 1,566 km, respectively and lowest for Zones 2 and 4 at 1,196 km and 1,119 km, respectively.

Not surprisingly, transport-related fuel consumption and CO₂ emissions also increase the further a residential development is located from the city centre. Figure 7 shows the annual transport-related CO₂ emissions per capita for each zone. It can be seen that annual transport emissions in Zone 4 were highest where the average person consumed 14.2 GJ of energy (equivalent to 449 l of petrol) and emitted 866 kg of CO₂ in meeting their 2006 transport needs. These figures decrease with increased proximity to the city centre and reach a minimum of 8.3 GJ and 472 kg CO₂ for Zone 1. Transport-related CO₂ emissions are therefore almost 85% higher for residential developments in the commuter towns than in the city centre and approximately 16% and 54% higher for suburban (Zone 2) and extra-urban (Zone 3) areas, respectively.

3.3 Operation 489

Results for the annual operational (heating, hot water and small power) energy requirements and associated CO₂ emissions for four different dwelling types are shown in Table 2 both on a per square metre and per dwelling basis. Detached houses were found to emit 88 kg CO₂/m²/annum of floor area; the corresponding figures for semi-detached houses, terraced houses and apartments were 67, 55 and 48 kg/m², respectively. Detached houses consume the greatest amounts of energy due to their higher floor areas and completely exposed external envelopes. Conversely, apartments have the lowest per dwelling consumption due to both their smaller size and reduced area of external envelope. Results were compared to emissions statistics for Scotland due to its similar weather and demographics; these are presented in Fig. 8. It can be seen that Scottish emissions compare favourably but are lower for detached and, to a lesser extent, semi-detached houses and are similar for terraced houses and apartments. These differences may be accounted for by the lower CO₂-intensity of electricity production in Scotland and the relatively large sizes of detached and semi-detached houses constructed in the GDA over the period being analysed.

Figure 7 shows annual operational CO₂ emissions per capita in each zone, taking account of the mix of house types and average occupancies for each zone. On average, each person in Zones 3 and 4 consumes approximately 85% more energy than in Zone 1 and 30% more than in Zone 2 in order to heat and power their home. This is due to the dominance of detached houses in the exurbs and commuter towns and the greater concentration of apartments in the city centre and suburbs. This effect of housing type distribution on increasing per capita operational energy use in Zones 3 and 4 is offset to some degree by higher dwelling occupancy rates in these areas (3.0 per dwelling versus 2.5 for urban).

3.4 Maintenance and demolition 525

The average energy requirements associated with maintaining and improving a domestic dwelling were estimated

Table 2 Annual operational energy requirements and associated CO₂ emissions for different dwelling types

Dwelling type	Energy requirements		CO ₂ output		Average size (m ²)
	(GJ/dwelling/year)	(MJ/m ² /year)	(t/dwelling/year)	(kg/m ² /year)	
Apartments	32	400	3.8	48	75
Terraced	65	623	5.8	55	105
Semi-Detached	92	763	8.1	67	121
Detached	161	886	15.9	88	181

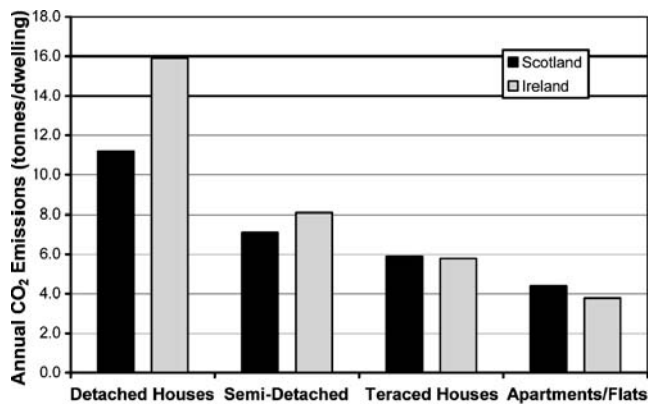


Fig. 8 CO₂ emissions per dwelling for Scottish housing stock and GDA sample (source: Hinchliffe 2005)

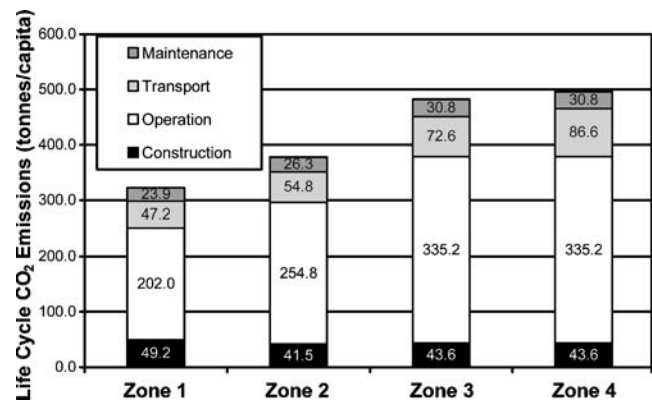


Fig. 9 Life cycle component CO₂ emissions for each zone (demolition has been excluded since emissions are negligible compared to other life cycle stages)

528 using input–output analysis (as described in Section 2.5) to
 529 be 0.046 GJ/m²/annum producing 6.8 kg CO₂/m²/annum.
 530 Due to a lack of data it was not possible to differentiate
 531 between maintenance requirements for different dwelling
 532 types. It is likely that this aggregation problem under-
 533 estimates maintenance energy requirements for detached
 534 and semi-detached houses given the greater opportunities
 535 for the major refurbishment and extension of such
 536 properties over apartments and—to a lesser extent—
 537 terraced houses. Therefore, in the absence of more
 538 disaggregated data, the same per square metre energy and
 539 CO₂ emissions intensities were applied to the average
 540 house size for each zone.

541 The cost of demolishing a typical 120 m² domestic
 542 dwelling (brick, block and timber floor construction) was
 543 estimated at € 17,500 in 2004 and the energy intensity for
 544 Nace 45.1 was calculated to be 2.25 MJ/€ of demolition
 545 expenditure. Combining these figures gives a housing
 546 demolition energy requirement of approximately 0.33 GJ/m²
 547 with associated emissions of 8.3 kg CO₂/m². The additional
 548 cost of demolishing the concrete or steel structure in an
 549 apartment is estimated to be approximately € 14,300,
 550 resulting in an apartment demolition energy intensity of
 551 0.60 GJ/m² with associated emissions of 15.1 kg CO₂/m².
 552 These figures do not allow for recycling.

553 3.5 Life cycle

554 Figure 9 shows the CO₂ emissions for a residence built in
 555 the last 10 years in each of the four zones in the GDA. This
 556 assumes a building lifespan of 100 years. On average Zone
 557 3 and 4 occupants consumed 92–98% more energy than
 558 those in Zone 1 and approximately one third more than
 559 Zone 2. Per capita CO₂ emissions in Zones 3 and 4 were
 560 approximately 50–55% greater than Zone 1 and almost
 561 33% greater than Zone 2. The higher CO₂/energy ratio in
 562 Zone 1 can be accounted for by the high incidence of CO₂-
 563 intensive electrical storage heating in Zone 1 apartments.

564 Figure 10 shows CO₂ emissions for each dwelling type
 565 excluding transport but including the construction,
 566 operation and maintenance life cycle stages (demolition
 567 has been excluded since these emissions are negligible).
 568 It can be seen that operational emissions dominate and
 569 that total emissions over 100 years are significantly
 570 higher for detached houses (1,773 t) than for semi-
 571 detached (929 t), terraced (680 t) and apartments (482 t).
 572 This is due to the greater electrical and space heating
 573 requirement associated with the large floor areas of
 574 detached houses in Ireland and their relatively greater
 575 exposed-envelope/volume ratios. These figures demon-
 576 strate significant benefits in promoting high-density
 577 spatial planning policies.

578 When the data for all zones are aggregated, operational
 579 requirements dominate accounting for 68% of all CO₂
 580 emissions followed by transport at 17%, construction at 9%
 581 and maintenance at 6%. Demolition contributes to less than
 582 0.5% of life cycle emissions.

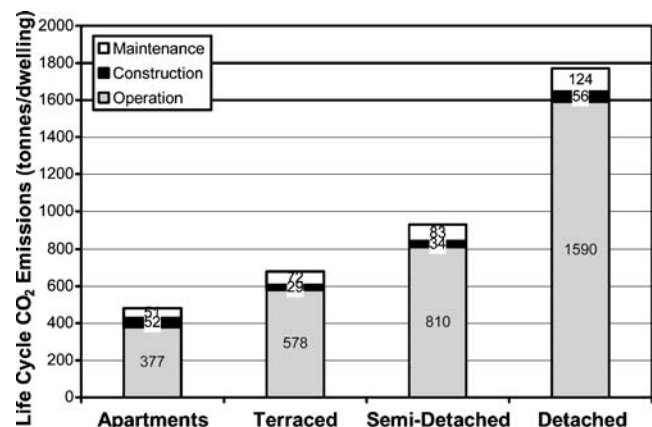


Fig. 10 CO₂ emissions for each dwelling type excluding transport (demolition has been excluded since emissions are negligible compared to other life cycle stages)

583 **4 Conclusions**

584 The life cycle per capita CO₂ emissions in the GDA are
 585 approximately 50–55% greater in the exurbs and commuter
 586 towns than in the city centre; corresponding energy
 587 intensities vary by almost 100%. Per capita residential
 588 emissions over a 100-year life span taking into account
 589 construction, operational, transport, maintenance and
 590 demolition energy requirements were 323, 378, 482 and
 591 496 tonnes, respectively for the urban, suburban, extra-
 592 urban and commuter town zones studied.

593 Of the five life cycle stages studied, operational energy
 594 requirements (predominantly space heating and hot water,
 595 but including power) contributed most significantly to
 596 emissions (68%), followed by transport (17%), construction
 597 (9%) and maintenance/renovation (6%); demolition was
 598 found to have a negligible effect. Annual per-capita
 599 operational emissions in Zones 3 and 4 were almost twice
 600 as great as those in Zone 1 both due to larger dwelling sizes
 601 and the predominance of detached and semi-detached
 602 dwellings (with large amounts of exposed walls) in the
 603 former and the predominance of smaller apartments in the
 604 latter. The car dominated as the preferred mode of transport
 605 in all zones; however, this was most pronounced in the
 606 zones furthest from the city centre (zones 3 and 4) where
 607 per capita emissions were almost twice those of residents in
 608 the city centre. It is interesting to note that despite their
 609 smaller size, the per capita construction energy intensities
 610 of apartments were approximately one third greater than for
 611 traditional detached, semi-detached and terraced residences
 612 due to the greater energy intensity of the structure.
 613 However, this difference was more than compensated for
 614 by the significantly lower operational emissions referred to
 615 above.

616 By 2006 residential development in the GDA completed
 617 over the 10 years up to and including that year was
 618 contributing 3,434 kt CO₂/annum—representing 4.9% of
 619 national emissions for that year; this comprised 2,108 tonnes
 620 of recurrent emissions (operational, transport, maintenance)
 621 and 1,325 kt of once off construction-related emissions.
 622 Had the development policy required Zone 1-type develop-
 623 ment and transport modes, then emissions for the year
 624 2006 would have been 2,892 kt CO₂—a reduction of
 625 almost 16% over the actual figure and representing
 626 approximately 4.1% of national emissions. However, in
 627 this scenario recurrent emissions would have been reduced
 628 to 1,417 kt CO₂—a reduction of 33% over actual levels.

629 **5 Recommendations and perspectives**

630 Irish and other governments' residential sector energy
 631 policies tend to focus on minimising operational energy

use and emissions. This approach is broadly supported by 632
 the findings of this paper that approximately two thirds of 633
 emissions arise from heating and powering dwellings in the 634
 GDA. However, although operational CO₂ emissions 635
 dominate those from transport, maintenance and construc- 636
 tion, assessing only operational emissions would underes- 637
 timate 100-year emissions by approximately one third. This 638
 demonstrates the benefits of LCA in evidence-based policy- 639
 making and highlights the need for policy measures targeted 640
 at mitigating non-operational emissions. Indeed, the relative 641
 importance of these emissions will increase as existing 642
 energy policies further reduce operational emissions. 643

644 An additional weakness of existing policies is the
 645 emphasis placed on reducing unit area emissions through
 646 energy efficiency measures. In Ireland, this policy has run
 647 concurrently with the development of large detached
 648 dwellings in the area studied. However, the results above
 649 clearly indicate that large, detached dwellings achieve
 650 national emissions policy objectives despite having a
 651 significantly higher environmental impact than smaller
 652 'attached' dwellings such as apartments. Although other
 653 social and functional considerations must be taken into
 654 account, a policy favouring high-density residential devel-
 655 opment with more modest floor areas would reduce
 656 emissions. Such a 'high density' planning approach would
 657 have the added benefit of reducing car dependency since
 658 residences could be constructed closer to urban centres and
 659 public transport infrastructure with relatively low emissions
 660 intensities. The operational and transport emissions from
 661 the worst performing zones could be reduced by approxi-
 662 mately 40% with suitable policies.

663
 664 **Acknowledgement** The author wishes to thank for the support of
 665 those who assisted in providing data for this work: the Pierser Group,
 666 Dublin Transportation Office, CODEMA, and Sustainable Energy
 667 Ireland. Funding for this work was provided by the Dublin Institute of
 668 Technology's Research Support Unit.

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